

# Response of *Fagus sylvatica* and *Picea abies* to the interactive effect of neighbor identity and enhanced CO<sub>2</sub> levels

V. Rolo<sup>1,2</sup> · E. Andivia<sup>1,3</sup> · R. Pokorný<sup>1,4</sup>

Received: 3 November 2014/Revised: 30 April 2015/Accepted: 11 May 2015/Published online: 19 May 2015  
© Springer-Verlag Berlin Heidelberg 2015

## Abstract

**Key message** Enhanced levels of CO<sub>2</sub> affected both the nutrition and morphology of both species. The effect of interspecific competition was dependent on the species identity but not on the CO<sub>2</sub> level.

**Abstract** The interest in adaptive forest strategies to overcome predicted scenarios of climate change is increasing worldwide. An example of these strategies is the introduction of native species into mono-specific plantations. However, to fully consider this option/strategy, a higher understanding of the responses of forest tree species to concurrent biotic and abiotic factors is needed. The aim of the present study was to assess nutritional and morphological adjustments of individuals of European beech (*Fagus sylvatica* L.) and Norway spruce [*Picea abies* (L.) Karst] growing at enhanced levels of CO<sub>2</sub> and with different proportions of con-specific individuals in its vicinity.

Individuals that grew at elevated CO<sub>2</sub> levels showed higher values of relative growth rate (RGR), total twig dry biomass and root biomass, and lower values of leaf area ratio, leaf N and Mg concentrations and soil nutrient concentrations. Individuals of Norway spruce growing in the vicinity of high proportions of European beech showed a reduction in the allocation of biomass to foliar tissue, and lower values of RGR and root biomass. European beech, by contrast, showed a limited response to Norway spruce presence and higher capacity in the exploitation of space both above- and below-ground. In conclusion, the lower response of European beech to both environmental factors suggests that the introduction of European beech into Norway spruce stands could be a feasible option in current forest transition strategies.

**Keywords** Mixed-forest · Climate change · Root morphology · Growth · Non-structural carbohydrates · CO<sub>2</sub> fumigation · Plant-to-plant interactions

Communicated by R. Matyssek.

**Electronic supplementary material** The online version of this article (doi:10.1007/s00468-015-1225-0) contains supplementary material, which is available to authorized users.

✉ V. Rolo  
victorroloromero@gmail.com

<sup>1</sup> Department of Silviculture, Faculty of Forestry and Wood Technology, Mendel University, Brno, Czech Republic

<sup>2</sup> Present Address: Conservation Ecology Research Unit, Faculty of Natural and Agricultural Sciences, University of Pretoria, Lynnwood Rd, Pretoria 0002, South Africa

<sup>3</sup> Present Address: Earth and Life Institute, Environmental Sciences, Université catholique de Louvain, Croix du sud 2, 1348 Louvain-la-Neuve, Belgium

<sup>4</sup> Global Change Research Centre AS CR, Belidla 4a, Brno 603 00, Czech Republic

## Introduction

Introducing autochthonous species into mono-specific plantations is a successful way to increase diversity, reduce natural damages and attenuate the risk of investments under predicted climatic scenarios (Zerbe 2002; Gärtner and Reif 2004; Knoke et al. 2005; Seidl et al. 2009). The awareness of developing adaptive forest management strategies is mainstream in Europe and worldwide (Bolte et al. 2009). However, contrasting results exist regarding whether mixed plantations can achieve greater productivity than monocultures (Pretzsch and Schütze 2009; Pretzsch et al. 2010). And the consideration of a lower profitability of mixed-forest than monocultures is widespread among

stakeholders and forest economists (Knoke et al. 2008). In this context, a better understanding of the responses of forest tree species to interspecific interactions in mixed-forest is necessary to predict their long term sustainability and test if mixed-forest can be a sustainable management option (Pretzsch et al. 2014).

The need to develop adaptive strategies is especially relevant for commercial mono-specific forest plantations of coniferous in Europe, where a severe reduction of forest stand's resistance against storm, snow, ice, drought and insect damage has been observed (Spiecker 2003). For example, mono-specific plantations of Norway spruce [*Picea abies* (L.) Karst], one of the most commercially valuable tree species in Europe, are declining (Jonard et al. 2012; Bošel'a et al. 2014). However, the survival of Norway spruce can be increased when growing in mixed-forest (Griess et al. 2012). The introduction of European beech (*Fagus sylvatica* L.) into coniferous mono-specific stands plays a central role in current forest transition strategies (Geßler et al. 2007). European beech is one of the most abundant and dominating tree species of the potential natural vegetation in temperate forests of Europe and several methods exist for the conversion of pure Norway spruce stands into mixed stands with European beech (Bravo-Oviedo et al. 2014). However, the long term sustainability of mixed stands depends on factors such as species identity or site conditions that can define the net outcome of mixing different species (Pretzsch et al. 2010). In addition, available information on how different species in mixed-forest responds to concurrent biotic interaction under predicted scenarios of climate change, such as the increase of atmospheric CO<sub>2</sub>, is much more limited (Geßler et al. 2007; IPCC 2007; Smith et al. 2013b).

Under elevated CO<sub>2</sub> conditions, tree species can increase plant growth directly by enhancing photosynthetic capacity or indirectly by increasing water or nutrient use efficiency (Körner 2006). However, the fertilization effect of CO<sub>2</sub> may be limited by the progressive scarcity of nutrient supplies, especially N (Reich et al. 2006; Dieleman et al. 2010) or by a lack of positive growth response despite the stimulation of the C assimilation rate induced by CO<sub>2</sub> (Körner 2006; Kirschbaum 2011). Additionally, in natural conditions, the response of an individual can depend on the identity of neighboring species (Lau et al. 2010) because processes affecting growth, such as nutrient acquisition, can be contingent on the identity of neighboring species under enhanced levels of CO<sub>2</sub> (Friend et al. 2000; Zak et al. 2012). These idiosyncratic responses to high CO<sub>2</sub> levels question the reasonable expectation that only superior competitors would be benefited in a mixture of tree species. Assessing how interspecific competition and high levels of CO<sub>2</sub> affects individual tree performance can help to

understand the response of mixed-forest under future climate scenarios (Niinemets 2010).

The net outcome of interspecific interactions between broadleaved and coniferous species, specifically between European beech and Norway spruce, under enhanced CO<sub>2</sub> levels is still largely unclear. At the sapling stage, Norway spruce can benefit from high CO<sub>2</sub> levels and outcompete European beech (Kozovits et al. 2005). But, the combined effect of CO<sub>2</sub> levels and other environmental factors, especially soil resources, can also mediate interspecific relations leading, for instance, European beech to benefit from high CO<sub>2</sub> levels and outcompete Norway spruce under certain circumstances (Spinnler et al. 2002). Early differences in growth or morphology between both species can determine their performance at the adult stage where differences in crown structure or in the efficiency to occupy space per structural cost can likely affect the resilience of the stand to disturbances (Reiter et al. 2005; Pretzsch 2014). Nevertheless, our understanding of how interspecific relations vary as the proportion of neighbors of a different species identity change is still scarce. Most studies analysing interspecific interactions between European beech and Norway spruce are based on 1:1 mixtures. This ratio seldom occurs outside controlled conditions. Thus, the application of knowledge acquired in 1:1 mixtures to natural conditions, where different mixture ratios are common, is limited. Testing different mixtures of European beech and Norway spruce may provide more comprehensive knowledge on the underlying mechanism that operates between both species and facilitate the generalization of the results of experiments under controlled conditions to more natural and realistic situations.

The aim of the present study was to assess how interacting relations of European beech and Norway spruce respond to CO<sub>2</sub> environment and if this effect varies when the proportion of neighboring individuals of a different identity change. We evaluated growth, biomass allocation to foliar or supporting tissue at the twig level, belowground morphological parameters and nutritional conditions in saplings of European beech and Norway spruce exposed to 6 years of CO<sub>2</sub> enrichment growing at different admixture treatments. In addition, we assessed soil nutrient concentrations to evaluate if the responses of the individuals to CO<sub>2</sub> environment and admixture treatments were modulated by this resource. Specifically, we aimed to answer the following questions: (1) is the individual response to CO<sub>2</sub> enrichment dependent on the proportion of neighboring species of a different identity growing in its vicinity or (2) are both pressures acting independently? (3) Are the observed responses contingent on the species identity? (4) Is there any mediating effect of soil nutrient concentration? We hypothesized that individuals of different species identity can respond differently to CO<sub>2</sub> environment, but

this effect may vary between different levels of admixture. Studying the response of individual plants to intra- and interspecific mixtures can be of key interest to better planning the management of mixed-forest where different levels of admixture occur.

## Materials and methods

### Study design and sampling scheme

The experiment was carried out in Beskydy Mountains (Bílý Kříž; 49°30'N, 18°32'E, 908 m a.s.l., NE of the Czech Republic). In this region, planted Norway spruce forests are the predominating vegetation type. The altitudinal range varies between 500 and 1700 m above sea level (a.s.l.). A moderately cold and wet mountain climate is typical for this region, annual average temperature of the study site is 4.9 °C and yearly accumulated rainfall is 1100 mm. Soils of the study site are characterized as ferric podzols. Soil depth is moderate and clay fraction range between 15 and 35 %. The geological bedrock is formed by Mesozoic Godula sandstone (flysch type).

The measurements were carried out in two adjoined glass domes (10 × 10 m in length and 7 m height in the central part), at ambient (385  $\mu\text{mol CO}_2 \text{ mol}^{-1}$ ) and elevated (700  $\mu\text{mol CO}_2 \text{ mol}^{-1}$ )  $\text{CO}_2$  concentrations. Glass domes resembled the well-known open-top chambers (OTCs) but that included a few modifications, such as adjustable windows (72 in total) made of safety glass that can be opened in 10° steps from fully close to fully open. The modifications were designed to overcome the problems of strong wind, humidity and temperature gradients associated with OTCs. On average, temperature and humidity gradients were maintained inside the domes at similar levels than natural conditions for more than 80 % of the time. In the glass dome with elevated  $\text{CO}_2$  concentrations,  $\text{CO}_2$  levels were maintained within 600–800  $\mu\text{mol CO}_2 \text{ mol}^{-1}$  ca. 72 % of the time during the growing

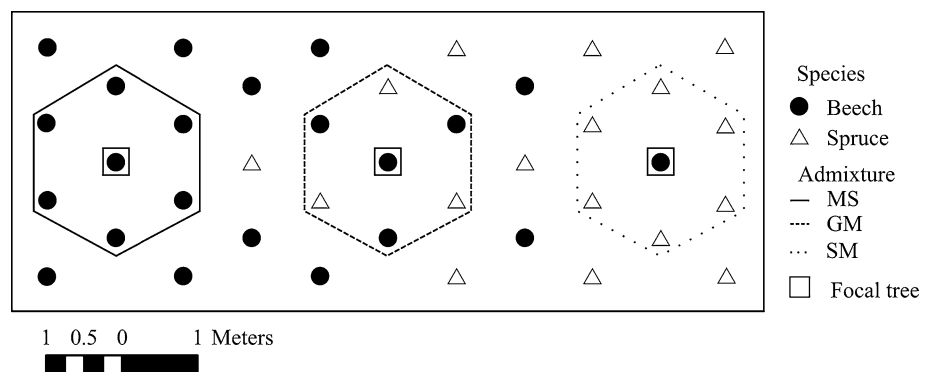
season. Soils inside the glass domes were not mechanically modified and are natural soils from the study zone. A priori differences in soil properties between the elevated and ambient glass domes were unlikely because of their close proximity to each other (less than 2 m). Further description of the chambers design and installation can be found in Urban et al. (2001).

A total of 99 seedlings per chamber were planted in 2006 in eleven lines with a separation of 1 m between plants and 0.85 m between lines. On average, the stem diameter at the time of planting of Norway spruce and European beech seedlings was  $10.1 \pm 0.26$  and  $11.3 \pm 0.45$  mm (mean  $\pm$  SE), respectively. Plant species were distributed to achieve the different admixtures treatments. There were three admixture treatments that were defined according to the number of neighbors belonging to the same species identity as the individual that was sampled (hereafter focal) (Fig. 1). Briefly, mono-specific (MS): all neighbors are individuals of the same species; group-mixture (GM): half of the neighbors are con-specific and the other half not; single-mixture (SM): no neighbors from the same species (Fig. 1). In summary, a total of 48 individuals were sampled (2 chambers  $\times$  2 species  $\times$  3 admixture treatments  $\times$  4 individuals). All sampled plants were surrounded by six individuals and no focal plants were located in the first lines close to the walls to avoid edge effects.

### Plant growth and components of biomass allocation at the twig level

Sapling stem diameter was measured yearly from 2006 to 2012 in each individual at the end of the growing season. Stem diameter was always measured in the same zone (30 cm height) by marking a line around the sapling stem. In 2012, once primary growth was fully achieved and during two consecutive days, current-year apical twigs were cut from four branches, one per orientation, to assess variation of aboveground biomass allocation among species and treatments. Branches were selected approximately at

**Fig. 1** Schematic representation of admixture treatment layout, *MS* mono-specific, *GM* group-mixture and *SM* single-mixture, per  $\text{CO}_2$  treatment showing European beech as a focal tree and Norway spruce as neighbor species. The distribution of Norway spruce was the same but it is not shown for simplicity



the same sapling height (1.3 m), corresponding to the upper third of the canopy in all cases. Twigs clearly located underneath other branches were not sampled to avoid confounding effects due to shadow.

From each twig, leaf/needle was removed and scanned (Epson Expression 10000XL) and projected leaf/needle area (LAp cm<sup>2</sup>) was measured with ImageJ (1.46r). All plant material was oven-dried (65 °C for 48 h) and weighed. Twig leaf/needle area ratio (LAR cm<sup>2</sup> g<sup>-1</sup>) was determined by dividing LAp by the total twig dry weight (TW g), including all leaf/needle and twig. Leaf/needle mass fraction (LMF %) was calculated by dividing total leaf/needle dry weight by total twig dry weight.

### Fine root sampling and processing

Soil samples were collected from the uppermost soil layer (15 cm) by means of a soil core (bipartite root auger Eijkelkamp, NL) in 2012, on the same days as twig sampling. One sampling point was placed randomly around each focal tree approximately half way between the stem and the drip line of each individual, but no closer than 20 cm to the stem. Soil samples were kept fresh at 4 °C until analysis. Root samples were separated from soil and washed using different filters (mesh size between 2 and 0.125 mm) to avoid fine roots losses. Soil was stored and dried for subsequent chemical analysis (see below). All roots less than or equal to 2 mm in diameter were classified as fine roots. Coarse roots and roots with symptoms of being dead (dark discoloration of the central cylinder and a decreased flexibility of root segments) were excluded from the analysis. Fine roots were identified visually and separated manually per species according to Schmid (2002). Briefly, *Picea* roots are elastic with a relatively thick and irregularly structured brownish cortex, whereas *Fagus* roots are less elastic and the red-brown cortex is thin with lines along the longitudinal axis. Root material was scanned (Epson Expression 10000XL) and images were processed with WinRhizo<sup>TM</sup> (Réagents Instruments Inc., Canada) to assess root morphological attributes: root length density (RLD cm/dm<sup>3</sup>) and average diameter (AD, mm). Subsequently, roots were oven-dried (65 °C for 48 h) and weighed. Differences in the abundance of both species inherent in our study design could preclude the direct comparison of fine root attributes between admixture treatments. Although focal trees were sampled under its canopy, a higher proportion of the opposite species identity in its vicinity may increase the probability of finding its roots in the sample. To avoid this bias, the standardization of total core values of Norway spruce and European beech to one unit of total basal area for the corresponding species surrounding the sample point has been showed to be an effective adjustment (Schmid 2002; Bolte and Villanueva 2006).

### Chemical analysis

Composite leaf/needle samples from the four twigs per individual were ground to pass a 250 µm mesh. Total nitrogen (N) and carbon (C) were measured by high-temperature (1000 °C) dry combustion method with an automatic analyser LECO CNS 2000. Nutrient concentrations (P, Mg, Ca and K) were measured with an ICP analyser (ICP-OES, Yobin Yvon Última 2, Tokyo, Japan) after wet-digestion with H<sub>2</sub>SO<sub>4</sub>/H<sub>2</sub>O<sub>2</sub> (Temminghoff and Houba 2004). Soil samples were sieved to pass a 2 mm mesh. Fine-earth was used to analyse total nitrogen and carbon with the same method used for leaves/needles, UV–visible spectrophotometry with Mehlich 3 extraction was used for P determination, and the Ca, Mg and K were determined by atomic absorption spectrophotometry with ammonium acetate extraction.

Leaf/needle and fine root non-structural carbohydrates concentrations (NSC) were determined colorimetrically by the anthrone method (Fales 1951) using glucose as standard. Ethanol-soluble fraction was extracted with 80 % ethanol at 30 °C during 30 min and the residue was boiled for 3 h at 100 °C with 3 % HCl to hydrolyze starch. Final values were found as the sum of ethanol-soluble carbohydrates and starch. NSC analyses were carried out in duplicate for each sample, further details on analyses and calculations can be found in Poorter and Villar (1997). Leaf NSC concentrations are contingent on the photosynthetic capacity of the leaves, as well as the amount of irradiance received (Lambers et al. 2008). This, together with the time delay that exists between photosynthate production and translocation, may affect the content of leaf/needle substances diurnally (Poorter et al. 2009). To partial out this source of variability, it has been recommended to express leaf/needle nutrient concentrations on a NSC-free dry weight (Poorter and Villar 1997; Niinemets 1997). Therefore, we expressed leaf/needle nutrient content on a NSC-free dry weight basis as:  $Nut_c = Nut / (1 - NSC)$ ; where Nut is the original nutrient concentration and NSC is expressed as the proportion of dry matter.

### Data analysis

To calculate sapling relative growth rate (RGR) we followed the recommendations of Paine et al. (2012). For each individual, we regressed basal diameter as a function of time and fitted a power-law function to the data. Depending on the value of the exponent, power-law models satisfactorily allow RGR to decrease as biomass increases, a fundamental property for growth analysis. We calculated RGR values per sapling and year from the solution of the power-law differential equation as a function of time (Paine et al. 2012). To analyse the effect of CO<sub>2</sub> treatment (A and E),

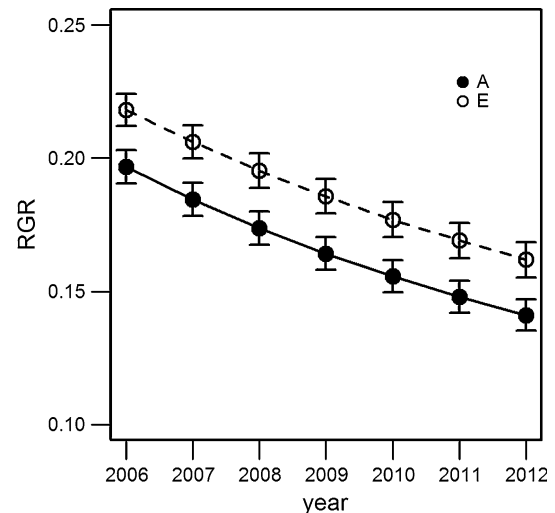
species identity (European beech and Norway spruce) and admixture treatment (MS, GM and SM) on the temporal dynamics of RGR, we employed linear mixed models. To take into account the temporal autocorrelation of the dependent variable, we included a temporal correlation structure in the model (Zuur et al. 2009) and sapling as a random effect to avoid pseudo-replication. The final set of explanatory factors and the best temporal correlation structure (AR-1) were selected according to the lowest AIC.

To test the effect of the independent variables ( $\text{CO}_2$ , species identity and admixture) and their interactions on multivariate data sets of leaf morphology (twig weight, leaf area ratio and leaf mass fraction), leaf nutrient concentration (N, P, K, Ca and Mg), root system (root biomass, root diameter and root length density) and soil nutrient content (N, P, K, Ca and Mg), we performed permutational multivariate ANOVA using Euclidean distance and 10,000 permutations. Multivariate datasets were projected onto the first two principal component axes to visually assess the significant differences found between treatments in the permutational multivariate ANOVA. Subsequently, we tested the effect of the independent variables on each univariate response variable by means of linear mixed models, including sapling as random factor. To account for differences in sapling size in the effect of twig morphological and root system variables, we included stem diameter at the time of sampling as a covariate in the models. The consideration of sapling stem diameter as a proxy of size allowed us to avoid its confounding effect when testing main factor effects. When a significant interaction effect was found, Tukey-HSD post hoc analyses were carried out. To explore the relationship per  $\text{CO}_2$  treatment between overall soil nutrient content and sapling growth, Mantel tests were carried out. Dependent variables were transformed when necessary to comply with normality and homogeneity of variance assumptions. Results are expressed as mean values  $\pm$  SE of the mean. All statistical analyses were performed in R v 3.0 (R Core Team 2013).

## Results

### Response of growth to admixture and $\text{CO}_2$ enrichment

Sapling growth showed independent responses to admixture and  $\text{CO}_2$ . Saplings showed higher values of RGR in the elevated environment of  $\text{CO}_2$  than in ambient ( $1.21 \pm 0.01$  and  $1.18 \pm 0.01$   $\text{cm cm}^{-1} \text{ year}^{-1}$ ;  $F = 4.1$ ,  $P = 0.049$ ) throughout the study period (Fig. 2), and this effect was consistent among admixture treatments and species (no significant interaction between  $\text{CO}_2$  and



**Fig. 2** Temporal dynamic of RGR based on stem diameter of European beech and Norway spruce growing at elevated (E) or ambient (A)  $\text{CO}_2$  treatments

admixture or  $\text{CO}_2$  and species). The effect of admixture on RGR was only apparent for Norway spruce (marginally significant interaction between species and admixture treatment,  $F = 2.9$ ,  $P = 0.062$ ). Indeed, the saplings of Norway spruce showed lower RGR values when growing in SM than in MS ( $1.17 \pm 0.01$  and  $1.24 \pm 0.01$   $\text{cm cm}^{-1} \text{ year}^{-1}$ ,  $P = 0.063$ ), while RGR values of European beech showed no response to admixture ( $1.18 \pm 0.01$  and  $1.17 \pm 0.01$   $\text{cm cm}^{-1} \text{ year}^{-1}$ ,  $P = 0.989$ ).

The relationship between soil nutrient concentration and RGR was dependent on the  $\text{CO}_2$  treatment. At ambient levels of  $\text{CO}_2$ , we observed a positive relationship between soil nutrient concentration and RGR (Mantel  $r = 0.13$ ,  $P = 0.071$ ), yet RGR was negatively related with soil nutrient concentration in the elevated  $\text{CO}_2$  environment (Mantel  $r = -0.21$ ,  $P = 0.004$ ).

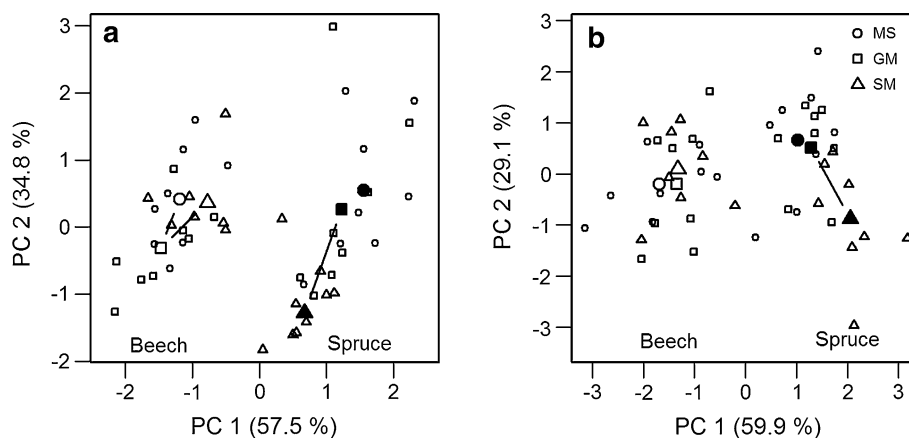
### Multivariate response of European beech and Norway spruce to $\text{CO}_2$ enrichment and admixture

The effect of  $\text{CO}_2$  treatment explained a high proportion of the multivariate variation across variable groups (Table 1). Soil nutrient concentration and root system attributes were significantly affected by  $\text{CO}_2$  treatment but their responses were dependent on species identity and admixture treatment, respectively. Leaf/needle nutrient concentration and twig morphology showed a lesser response to  $\text{CO}_2$  environment, with admixture treatment explaining a higher portion of the variance, but the effect of admixture seemed to depend on the species identity of the focal sapling (Table 1). The projection of twig morphology and leaf/needle nutrient concentration on a reduced space suggested

**Table 1** Results from perMANOVA testing the effect of CO<sub>2</sub> (E), Species (Sp), Admixture (Ad) and their interactions on multivariate datasets of leaf morphology (twig weight, leaf area ratio and leaf mass fraction), leaf nutrient concentration (N, P, K, Ca and Mg), root system attributes (root biomass, root diameter and root length density) and soil nutrient content (N, P, K, Ca and Mg)

	CO <sub>2</sub> (E)	Species (Sp)	Admixture (Ad)	E:Sp	E:Ad	Sp:Ad	E:Sp:Ad	Residual
Leaf morphology								
<i>R</i> <sup>2</sup>	0.008	0.863	0.009	0.001	0.005	0.036	0.0	0.074
<i>F</i>	4.0	419.0	2.3	0.7	1.3	8.7	0.2	
<i>P</i>	<b>0.054</b>	<b>&lt;0.001</b>	0.106	0.401	0.270	<b>&lt;0.001</b>	0.786	
Leaf nutrients								
<i>R</i> <sup>2</sup>	0.023	0.455	0.080	0.002	0.023	0.099	0.027	0.287
<i>F</i>	2.9	56.0	5.0	0.2	1.4	6.2	1.6	
<i>P</i>	<b>0.067</b>	<b>&lt;0.001</b>	<b>0.003</b>	0.747	0.216	<b>0.001</b>	0.173	
Root system								
<i>R</i> <sup>2</sup>	0.067	0.033	0.321	0.004	0.007	0.032	0.099	0.433
<i>F</i>	5.6	2.7	13.0	0.3	0.3	1.3	4.1	
<i>P</i>	<b>0.022</b>	<b>0.099</b>	<b>&lt;0.001</b>	0.539	0.742	0.271	<b>0.020</b>	
Soil nutrients								
<i>R</i> <sup>2</sup>	0.307	0.053	0.021	0.057	0.028	0.017	0.022	0.491
<i>F</i>	22.	3.9	0.7	4.2	1.0	0.6	0.8	
<i>P</i>	<b>&lt;0.001</b>	<b>0.052</b>	0.470	<b>0.047</b>	0.368	0.542	0.457	

Significant differences ( $P < 0.05$ ) are depicted in bold font and marginal significant in bold and italic ( $0.05 < P < 0.10$ )



**Fig. 3** Projected values of leaf morphology (a) and leaf nutrient concentration (b) multivariate dataset onto the first two principal components showing the interacting term between species (European beech and Norway spruce) and admixture, *MS* mono-specific, *GM*

group-mixture and *SM* single-mixture. The explained variance per axis is shown. *Open symbols* represent each sapling per admixture treatment. *Filled symbols* represent the centroid per admixture treatment connected by lines

that Norway spruce was more affected by admixture than European beech (Fig. 3). Indeed, we observed larger differences in the morphology of Norway spruce twigs and leaf/needle nutrient concentrations between *MS* and *SM* admixture treatments (larger separation between centroids) than in European beech (Fig. 3).

### Response of European beech and Norway spruce to CO<sub>2</sub> enrichment

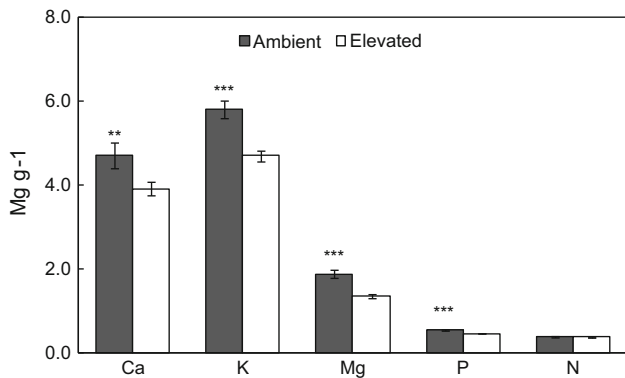
Further analysis of each univariate variable separately confirmed the effects of high CO<sub>2</sub> levels observed across multivariate datasets (Table S1 and S2). Saplings growing

in the elevated CO<sub>2</sub> environment had higher values of root biomass ( $3.25 \pm 0.66$  and  $2.17 \pm 0.54$  g m<sup>-2</sup> cm<sup>-2</sup> BA;  $F = 4.9$ ,  $P = 0.031$ ), total twig biomass ( $1.06 \pm 0.10$  and  $0.85 \pm 0.06$  g;  $F = 4.5$ ,  $P = 0.039$ ) and leaf/needle P concentration than in ambient (Table 2). Root length density values of European beech also showed a significant response to CO<sub>2</sub> environment, showing higher values in the elevated environment than in ambient ( $0.76 \pm 0.17$  and  $0.32 \pm 0.05$  cm dm<sup>-3</sup> cm<sup>-2</sup> BA;  $F = 3.4$ ,  $P = 0.003$ ). On the other hand, we observed the opposite response to CO<sub>2</sub> levels, with significant lower values in the elevated environment than in ambient, for leaf/needle area ratio ( $108.72 \pm 12.14$  and  $119.01 \pm 11.4$  cm<sup>2</sup> g<sup>-1</sup>;  $F = 4.3$ ,

**Table 2** Needles and leaves nutrient concentration (mg/g) ( $\pm$ SE) per CO<sub>2</sub> treatment

	Leaves/needles		<i>F</i>	<i>P</i>
	Ambient	Elevated		
N	19.30 $\pm$ 0.33	18.36 $\pm$ 0.43	2.8	<b><i>0.098</i></b>
P	2.09 $\pm$ 0.12	2.31 $\pm$ 0.09	3.9	<b><i>0.051</i></b>
K	8.78 $\pm$ 0.64	8.2 $\pm$ 0.65	2.2	0.143
Mg	1.84 $\pm$ 0.10	1.7 $\pm$ 0.10	3.2	<b><i>0.082</i></b>
Ca	7.65 $\pm$ 0.40	7.73 $\pm$ 0.56	0.1	0.892
C (%)	50.52 $\pm$ 2.33	50.17 $\pm$ 2.34	2.0	0.165
C:N	26.37 $\pm$ 0.51	27.96 $\pm$ 0.69	2.3	0.136
C:P	261.85 $\pm$ 16.4	228.08 $\pm$ 9.41	6.1	<b><i>0.018</i></b>
N:P	10.03 $\pm$ 0.69	8.18 $\pm$ 0.33	12.1	<b><i>0.001</i></b>

Leaf/needle nutrient concentrations are corrected values taking into account NSC content. Significant values are depicted in bold ( $P < 0.05$ ) and marginally significant in bold and italics ( $0.05 < P < 0.1$ )



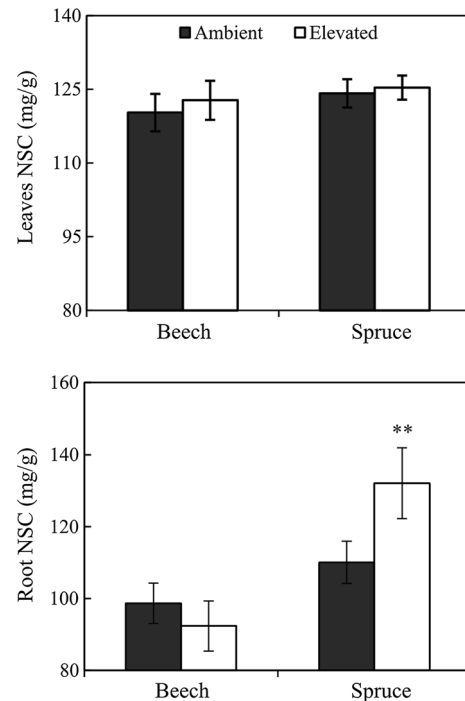
**Fig. 4** Soil nutrient concentration (mg/g) mean values ( $\pm$ SE) per CO<sub>2</sub> treatment, ambient and elevated. Asterisks depict significant differences at: \* $0.05 < P < 0.1$ , \*\* $P < 0.05$  and \*\*\* $P < 0.01$

$P = 0.045$ ), soil nutrient concentration, that was consistently lower for most of the nutrients analysed (Fig. 4), and leaf/needle N and Mg concentration (Table 2).

Higher levels of CO<sub>2</sub> did not affect leaf/needle C or NSC concentration. Norway spruce saplings showed higher root NSC concentrations in the elevated environment than in ambient, whereas European beech did not show any response (significant interaction between CO<sub>2</sub> treatment and species,  $F = 5.7$ ,  $P = 0.002$ ) (Fig. 5). The response of leaf/needle nutrient stoichiometric to CO<sub>2</sub> environment was also significant, showing lower values of C:P and N:P ratios at high levels of CO<sub>2</sub> than in ambient (Table 2).

**Response of European beech and Norway spruce to admixture**

The effect of admixture was mostly limited to twig morphological and root system attributes. In general, Norway



**Fig. 5** Leaves/needles and roots non-structural carbohydrates concentration (NSC, mg/g) mean values ( $\pm$ SE) of European beech and Norway spruce per CO<sub>2</sub> treatments, ambient and elevated. Asterisks depict significant differences at  $P < 0.05$

spruce showed a higher response to the presence of European beech than European beech to the presence of Norway spruce (Table 3). At the twig level, Norway spruce saplings showed significant lower values of total twig biomass and significant higher values of needle mass fraction and needle area ratio in SM than in MS. On the contrary, European beech showed significant lower values of leaf area ratio in SM than in MS (Table 3). Both species showed significant higher values of root biomass in MS than in SM. Norway spruce showed a significant reduction of root average diameter and no effect on root length density between MS and SM. By contrast, European beech showed a significant reduction in root length density and no effect on average diameter between both treatments.

**Discussion**

**Response of plant growth and twig morphological attributes to CO<sub>2</sub> and admixture**

Enhanced levels of CO<sub>2</sub> and interspecific competition modified the allocation of resources in European beech and Norway spruce; however, the amount of change depended either on the CO<sub>2</sub> environment or on the admixture treatment. According to our results, the response of each

**Table 3** Values (mean  $\pm$  SE) of twig morphological, total twig weight (g), leaf mass fraction (%), leaf area ratio ( $\text{cm}^2/\text{g}$ ), and root system, average diameter (mm), biomass ( $\text{g}/\text{m}^2 \text{ cm}^2 \text{ BA}$ ), length density ( $\text{cm}/\text{dm}^3 \text{ cm}^2 \text{ BA}$ ), attributes of European beech and Norway spruce, growing at different regimes of admixture, mono-specific (MS), group-mixture (GM) and single-mixture (SM)

	Species	Admixture			Interaction	
		MS	GM	SM	F	P
Twig morphology						
Total dry weight	Beech	1.24 $\pm$ 0.10	0.96 $\pm$ 0.1	1.14 $\pm$ 0.1	3.4	0.043
	Spruce	1.03 $\pm$ 0.14a	0.97 $\pm$ 0.19a	0.40 $\pm$ 0.05b		
Leaf mass fraction	Beech	0.72 $\pm$ 0.01	0.75 $\pm$ 0.02	0.69 $\pm$ 0.02	5.2	0.009
	Spruce	0.59 $\pm$ 0.02b	0.62 $\pm$ 0.02b	0.68 $\pm$ 0.01a		
Leaf area ratio	Beech	165.2 $\pm$ 7.9ab	178.68 $\pm$ 9.2b	155.9 $\pm$ 6.0a	6.5	0.004
	Spruce	46.66 $\pm$ 3.21b	55.92 $\pm$ 4.32b	80.82 $\pm$ 6.14a		
Root system						
Average diameter	Beech	0.29 $\pm$ 0.03	0.28 $\pm$ 0.01	0.28 $\pm$ 0.02	4.6	0.015
	Spruce	0.44 $\pm$ 0.03a	0.38 $\pm$ 0.03a	0.29 $\pm$ 0.02b		
Root length density	Beech	0.86 $\pm$ 0.23a	0.37 $\pm$ 0.08b	0.39 $\pm$ 0.13b	3.8	0.006
	Spruce	0.33 $\pm$ 0.05	0.26 $\pm$ 0.06	0.28 $\pm$ 0.06		
Root biomass	Beech	4.66 $\pm$ 0.81a	1.95 $\pm$ 0.46b	2.61 $\pm$ 0.65ab	5.9	<0.001
	Spruce	4.13 $\pm$ 0.84a	2.02 $\pm$ 0.48b	0.89 $\pm$ 0.23b		

Different letters per species and row indicated significant differences at  $P < 0.05$

species type to admixture would be similar at both levels of  $\text{CO}_2$ , enhanced and ambient. In other words, an individual of Norway spruce diminished its growth when growing in the vicinity of high proportions of European beech and this effect persisted under elevated  $\text{CO}_2$  conditions. This result contradicted our expectations because Norway spruce has been commonly regarded as a superior competitor than European beech (Matyssek et al. 2005) and examples of Norway spruce outcompeting European beech at elevated  $\text{CO}_2$  levels have been reported (Kozovits et al. 2005). The highest response of Norway spruce was observed when growing in high densities of European beech, suggesting that shading could have limited Norway spruce growth (Kirschbaum 2011). Acclimation of plants to shade involves higher allocation to leaf area per gram plant, among other adjustments. The higher values of leaf/needle area ratio or leaf/needle mass fraction in SM than in MS observed in Norway spruce, but not in European beech that showed the opposite pattern (Table 3), supports the notion of shading induced by European beech. The tendency to conserve resources in plants growing in shade, could explain the lower RGR values observed in Norway spruce than in European beech.

Results of European beech outcompeting Norway spruce have only been shown in nutrient rich calcareous soils (Spinnler et al. 2002). Interestingly, the soils of our studied zone are mostly acidic ( $\text{pH} \sim 4$ ) suggesting that other factors may play a major role in determining the direction of this interspecific interaction. These discrepancies may arise due to the different admixture treatments employed in both studies. The lack of differences observed in our study between MS and GM admixture treatments

suggests that experiments assessing the effect of admixture at 1:1 proportions may not fully disentangle the effect of biotic interactions and be more dependent on external factors, such as soil type.

The adjustments of both tree species to admixture seemed to conflict with the overall effect of  $\text{CO}_2$  enrichment on aboveground resources allocation, especially in Norway spruce. While admixture tended to optimize light harvesting by increasing the allocation of resources to leaves,  $\text{CO}_2$  fumigation tended to induce the opposite. We observed that both species showed lower values of leaf area ratio when growing at enhanced levels of  $\text{CO}_2$ . A proportionally higher reduction in the amount of photosynthetically active tissues and a higher sequestration into structural and conducting tissues may be related to processes of downward regulation (Poorter and Navas 2003). This result would be in accordance with the observed reduction in N and Mg concentration at high  $\text{CO}_2$  levels and would point to a diminution of the photosynthetic capacity of the studied trees (Lukac et al. 2010). Conflicting requirements for biomass investment in different plant organs to increase the tolerance to different stresses are common in nature and may lead to certain cost and risk that may compromise plant growth and survival (Valladares and Niinemets 2008). Processes of down-regulation are often attributed to imbalances between source–sink relationships (Urban et al. 2003). However, we did not observe differences in the accumulation of NSC on leaf/needle between  $\text{CO}_2$  treatments (Fig. 5), suggesting a weak downward regulation effect in the leaf/needle cohort studied (Tissue et al. 2001) and limited consequences of the conflicting effect between environmental stresses.

## Response of root system attributes to CO<sub>2</sub> and admixture

The increase in root biomass in both species at high levels of CO<sub>2</sub> suggests a shift in the allocation of carbon to the root system (Norby et al. 2002). This flux could be especially relevant in Norway spruce which showed a significant increase in the concentration of NSC in roots. Indeed, we observed a significant correlation between root biomass and NSC concentration for Norway spruce ( $r = 0.71$ ,  $P = 0.009$ ), whereas we did not observe this effect in European beech ( $r = 0.52$ ,  $P = 0.175$ ). High root NSC concentrations in trees exposed to high levels of CO<sub>2</sub> can be interpreted as an accumulation of substrate available to support the production of fine roots (Norby et al. 2002). However, both species increased their fine root biomass at elevated levels of CO<sub>2</sub> similarly, suggesting that these discrepancies could be related to a different strategy of biomass allocation to roots between species. Similar to the aboveground compartment, European beech tended to use a higher volume of soil, especially at elevated levels of CO<sub>2</sub>. Indeed, European beech accompanied the increase in biomass by an increase in root length density, suggesting a higher use of belowground space under competitive environments (Grams et al. 2002; Bolte and Villanueva 2006). The effect of admixture on root attributes also support the notion of a different belowground strategy between both species. This effect together with the discrepancies observed at the twig level, could be an additional evidence of the negative consequences of admixing Norway spruce trees within a group of European beech at early developmental stages (Bolte et al. 2014).

## Effect of CO<sub>2</sub> and admixture on nutrient relations

High levels of CO<sub>2</sub> tended to modify the nutrient relations of both species. The consistent lower values of soil nutrient concentration in the elevated CO<sub>2</sub> environment, together with its negative association with RGR, suggest an increased difficulty in the access to soil resources (e.g. by a diminution in nutrient availability) which may have elicited an increased soil exploration by fine roots (Norby and Iversen 2006; Smith et al. 2013a). Extensive root systems exploring large volumes of soils are associated with the plant ability to capture less mobile elements, such as P, but not necessarily with more mobile nutrients, such as N (Sands and Mulligan 1990). In addition, the allocation of resources to the root system is not necessarily associated with a benefited nutrition and it might exacerbate the limitation of nutrients by enhancing the sink effect of other compartments, such as ectomycorrhizas and mycelia (Cudlin et al. 2007). As a result, the observed higher allocation of resources to the belowground compartment did not

guarantee a positive effect in leaf/needle nutrient concentrations, with the exception of phosphorus (Table 2). The positive effect observed in P nutrition may be related with a higher exudation of phosphatases which are related with P uptake and have been shown to be favored under CO<sub>2</sub> elevated environments (Duval et al. 2012). Nutrient concentrations values are related to the quantity of nutrients still embedded in organic molecules or bound to the soil matrix, thus the significant lower values observed suggest that plants tried to increase their nutrient uptake not only by increasing the amount of explored soil but also by triggering additional mechanisms to scavenge nutrients (Phillips et al. 2011). Nevertheless, despite the likely efforts to increase nutrient uptake, we observed a significant reduction in the N:P ratio in the elevated environment. Imbalances in nutrient stoichiometry of plant tissues are related to nutrient limitations and can have important implications for plant interactions with nutrient cycling (Elser et al. 2000; Güsewell 2004). Thus, our results suggest that long term fumigation may have an indirect effect on plant nutrition through the modification of litter quality, microbial community and the recycling of nutrients within the ecosystem.

## Conclusions

The morphology and nutrition of European beech and Norway spruce saplings showed a significant response to elevated levels of CO<sub>2</sub>. After 6 years of CO<sub>2</sub> enrichment, both species showed higher relative growth rate at elevated than ambient CO<sub>2</sub> levels, but the lower values of soil nutrient concentrations together with the higher allocation of resources to the root systems suggest a likely nutrient limitation effect in the elevated environment. The effect of admixture was dependent on the species identity and mostly limited to morphological attributes. Our results suggest that experiments assessing the effects of species mixture at 1:1 ratios may not be as easily transferable to natural conditions as previously supposed. We observed that species mixture at 1:1 ratios overlooked inter-specific responses that only occurred when a focal individual was outcompeted by many individuals of a different species, as commonly found in natural conditions. Overall, European beech seemed to cope better with both environmental pressures suggesting that the introduction of single individuals of this tree species into Norway spruce stands could be a feasible option in current forest transition strategies, at least at the sapling stage.

**Author contribution statement** VR and EA designed the sampling scheme, collected and processed the samples and wrote the manuscript. RP revised the manuscript.

**Acknowledgments** We thank Pavel Formánek for help and support with the laboratory analysis and to Czech Globe staff for allowing us

to use their installations and scientific facilities. This study was funded by the Ministry of Education, Youth and Sports of CR within the National Sustainability Program I (NPU I), grant number LO1415. VR and EA were supported by a postdoctoral grant from the OP Education for Competitiveness (European Social Fund and Czech Republic Ministry of Education, Youth and Sport CZ.1.07/2.3.00/30.0017). VR was also supported by the government of South Africa (NRF Freestanding Post-doctoral Fellowship) and EA by a FSR Incoming Post-doctoral Fellowship of the Académie Universitaire “Louvain” and the European Commission.

**Conflict of interest** The authors declare that they have no conflict of interest.

## References

- Bolte A, Villanueva I (2006) Interspecific competition impacts on the morphology and distribution of fine roots in European beech (*Fagus sylvatica* L.) and Norway spruce (*Picea abies* (L.) Karst.). *Eur J For Res* 125:15–26
- Bolte A, Ammer C, Löf M et al (2009) Adaptive forest management: a prerequisite for sustainable forestry in the face of climate change. In: Spathelf P (ed) Sustainable forest management in a changing world. Springer, Netherlands, pp 115–139
- Bolte A, Hilbrig L, Grundmann BM, Roloff A (2014) Understory dynamics after disturbance accelerate succession from spruce to beech-dominated forest—the Siggaboda case study. *Ann For Sci* 71:139–147. doi:10.1007/s13595-013-0283-y
- Boštel' a M, Sedmák R, Sedmáková D et al (2014) Temporal shifts of climate—growth relationships of Norway spruce as an indicator of health decline in the Beskids, Slovakia. *For Ecol Manag* 325:108–117. doi:10.1016/j.foreco.2014.03.055
- Bravo-Oviedo A, Pretzsch H, Ammer C et al (2014) European mixed forests: definition and research perspectives. *For Syst* 23:518–533
- Cudlin P, Kieliszewska-Rokicka B, Rudawska M et al (2007) Fine roots and ectomycorrhizas as indicators of environmental change. *Plant Biosyst* 141:406–425. doi:10.1080/11263500701626028
- Dieleman WIJ, Luyssaert S, Rey A et al (2010) Soil [N] modulates soil C cycling in CO<sub>2</sub>-fumigated tree stands: a meta-analysis. *Plant Cell Environ* 33:2001–2011
- Duval BD, Blankinship JC, Dijkstra P, Hungate BA (2012) CO<sub>2</sub> effects on plant nutrient concentration depend on plant functional group and available nitrogen: a meta-analysis. *Plant Ecol* 213:505–521. doi:10.1007/s11258-011-9998-8
- Elser JJ, Fagan WF, Denno RF et al (2000) Nutritional constraints in terrestrial and freshwater food webs. *Nature* 408:578–580. doi:10.1038/35046058
- Fales FW (1951) The assimilation and degradation of carbohydrates by yeast cells. *J Biol Chem* 193:113–124
- Friend AL, Jifon JL, Berrang PC et al (2000) Elevated atmospheric CO<sub>2</sub> and species mixture alter N acquisition of trees in stand microcosms. *Can J For Res* 30:827–836. doi:10.1139/x00-019
- Gärtner S, Reif A (2004) The impact of forest transformation on stand structure and ground vegetation in the southern Black Forest, Germany. *Plant Soil* 264:35–51. doi:10.1023/B:PLSO.0000047751.25915.77
- Geßler A, Keitel C, Kreuzwieser J et al (2007) Potential risks for European beech (*Fagus sylvatica* L.) in a changing climate. *Trees Struct Funct* 21:1–11. doi:10.1007/s00468-006-0107-x
- Grams TEE, Kozovits AR, Reiter IM et al (2002) Quantifying competitiveness in woody Plants. *Plant Biol* 4:153–158. doi:10.1055/s-2002-25729
- Griess VC, Acevedo R, Härtl F et al (2012) Does mixing tree species enhance stand resistance against natural hazards? A case study for spruce. *For Ecol Manag* 267:284–296. doi:10.1016/j.foreco.2011.11.035
- Güsewell S (2004) N: P ratios in terrestrial plants: variation and functional significance. *New Phytol* 164:243–266. doi:10.1111/j.1469-8137.2004.01192.x
- IPCC (2007) Climate change 2007: synthesis report. In: Core Writing Team, Pachauri RK, Reisinger A (eds) Contribution of working groups I, II and III to the fourth assessment report of the Intergovernmental Panel on Climate Change. IPCC, Geneva, Switzerland, p 104
- Jonard M, Legout A, Nicolas M et al (2012) Deterioration of Norway spruce vitality despite a sharp decline in acid deposition: a long-term integrated perspective. *Glob Change Biol* 18:711–725. doi:10.1111/j.1365-2486.2011.02550.x
- Kirschbaum MUF (2011) Does enhanced photosynthesis enhance growth? Lessons learned from CO<sub>2</sub> enrichment studies. *Plant Physiol* 155:117–124. doi:10.1104/pp.110.166819
- Knocke T, Stimm B, Ammer C, Moog M (2005) Mixed forests reconsidered: a forest economics contribution on an ecological concept. *For Ecol Manag* 213:102–116. doi:10.1016/j.foreco.2005.03.043
- Knocke T, Ammer C, Stimm B, Mosandl R (2008) Admixing broadleaved to coniferous tree species: a review on yield, ecological stability and economics. *Eur J For Res* 127:89–101. doi:10.1007/s10342-007-0186-2
- Körner C (2006) Plant CO<sub>2</sub> responses: an issue of definition, time and resource supply. *New Phytol* 172:393–411. doi:10.1111/j.1469-8137.2006.01886.x
- Kozovits AR, Matyssek R, Blaschke H et al (2005) Competition increasingly dominates the responsiveness of juvenile beech and spruce to elevated CO<sub>2</sub> and/or O<sub>3</sub> concentrations throughout two subsequent growing seasons. *Glob Change Biol* 11:1387–1401. doi:10.1111/j.1365-2486.2005.00993.x
- Lambers H, Chapin FS, Pons TL (2008) Mineral nutrition. *Plant physiology ecology*. Springer, New York, pp 255–320
- Lau JA, Shaw RG, Reich PB, Tiffin P (2010) Species interactions in a changing environment: elevated CO<sub>2</sub> alters the ecological and potential evolutionary consequences of competition. *Evol Ecol Res* 12:435–455
- Lukac M, Calfapietra C, Lagomarsino A, Loreto F (2010) Global climate change and tree nutrition: effects of elevated CO<sub>2</sub> and temperature. *Tree Physiol* 30:1209–1220
- Matyssek R, Agerer R, Ernst D et al (2005) The plant's capacity in regulating resource demand. *Plant Biol* 7:560–580. doi:10.1055/s-2005-872981
- Niinemets Ü (1997) Distribution patterns of foliar carbon and nitrogen as affected by tree dimensions and relative light conditions in the canopy of *Picea abies*. *Trees Struct Funct* 11:144–154. doi:10.1007/PL00009663
- Niinemets Ü (2010) Responses of forest trees to single and multiple environmental stresses from seedlings to mature plants: past stress history, stress interactions, tolerance and acclimation. *For Ecol Manag* 260:1623–1639. doi:10.1016/j.foreco.2010.07.054
- Norby RJ, Iversen CM (2006) Nitrogen uptake, distribution, turnover and efficiency of use in a CO<sub>2</sub>-enriched sweetgum forest. *Ecology* 87:5–14. doi:10.1890/04-1950
- Norby RJ, Hanson PJ, O'Neill EG et al (2002) Net primary productivity of a CO<sub>2</sub>-enriched deciduous forest and the implications for carbon storage. *Ecol Appl* 12:1261–1266
- Paine CET, Marthews TR, Vogt DR et al (2012) How to fit nonlinear plant growth models and calculate growth rates: an update for ecologists. *Methods Ecol Evol* 3:245–256. doi:10.1111/j.2041-210X.2011.00155.x

- Phillips RP, Finzi AC, Bernhardt ES (2011) Enhanced root exudation induces microbial feedbacks to N cycling in a pine forest under long-term CO<sub>2</sub> fumigation. *Ecol Lett* 14:187–194
- Poorter H, Navas M-L (2003) Plant growth and competition at elevated CO<sub>2</sub>: on winners, losers and functional groups. *New Phytol* 157:175–198. doi:10.1046/j.1469-8137.2003.00680.x
- Poorter H, Villar R (1997) The fate of acquired carbon in plants: chemical composition and construction costs. In: Bazzaz FA, Grace J (eds) *Plant resource allocation*. Academic Press, San Diego, pp 39–72
- Poorter H, Niinemets Ü, Poorter L et al (2009) Causes and consequences of variation in leaf mass per area (LMA): a meta-analysis. *New Phytol* 182:565–588. doi:10.1111/j.1469-8137.2009.02830.x
- Pretzsch H (2014) Canopy space filling and tree crown morphology in mixed-species stands compared with monocultures. *For Ecol Manag* 327:251–264. doi:10.1016/j.foreco.2014.04.027
- Pretzsch H, Schütze G (2009) Transgressive overyielding in mixed compared with pure stands of Norway spruce and European beech in Central Europe: evidence on stand level and explanation on individual tree level. *Eur J For Res* 128:183–204. doi:10.1007/s10342-008-0215-9
- Pretzsch H, Block J, Dieler J et al (2010) Comparison between the productivity of pure and mixed stands of Norway spruce and European beech along an ecological gradient. *Ann For Sci* 67:712
- Pretzsch H, Rötzer T, Matyssek R et al (2014) Mixed Norway spruce (*Picea abies* (L.) Karst) and European beech (*Fagus sylvatica* (L.)) stands under drought: from reaction pattern to mechanism. *Trees Struct Funct* 28:1305–1321. doi:10.1007/s00468-014-1035-9
- R Core Team (2013) R: a language and environment for statistical computing. R Foundation for Statistical Computing, Vienna
- Reich PB, Hungate BA, Luo Y (2006) Carbon–Nitrogen interactions in terrestrial ecosystems in response to rising atmospheric carbon dioxide. *Annu Rev Ecol Evol Syst* 37:611–636. doi:10.1146/annurev.ecolsys.37.091305.110039
- Reiter IM, Häberle K-H, Nunn AJ et al (2005) Competitive strategies in adult beech and spruce: space-related foliar carbon investment versus carbon gain. *Oecologia* 146:337–349. doi:10.1007/s00442-005-0146-9
- Sands R, Mulligan DR (1990) Water and nutrient dynamics and tree growth. *For Ecol Manag* 30:91–111. doi:10.1016/0378-1127(90)90129-Y
- Schmid I (2002) The influence of soil type and interspecific competition on the fine root system of Norway spruce and European beech. *Basic Appl Ecol* 3:339–346
- Seidl R, Schelhaas M-J, Lindner M, Lexer M (2009) Modelling bark beetle disturbances in a large scale forest scenario model to assess climate change impacts and evaluate adaptive management strategies. *Reg Environ Change* 9:101–119. doi:10.1007/s10113-008-0068-2
- Smith AR, Lukac M, Bambrick M et al (2013a) Tree species diversity interacts with elevated CO<sub>2</sub> to induce a greater root system response. *Glob Change Biol* 19:217–228. doi:10.1111/gcb.12039
- Smith AR, Lukac M, Hood R et al (2013b) Elevated CO<sub>2</sub> enrichment induces a differential biomass response in a mixed species temperate forest plantation. *New Phytol* 198:156–168. doi:10.1111/nph.12136
- Spiecker H (2003) Silvicultural management in maintaining biodiversity and resistance of forests in Europe—temperate zone. *J Environ Manage* 67:55–65
- Spinnler D, Egli P, Körner C (2002) Four-year growth dynamics of beech-spruce model ecosystems under CO<sub>2</sub> enrichment on two different forest soils. *Trees Struct Funct* 16:423–436. doi:10.1007/s00468-002-0179-1
- Temminghoff EEJM, Houba VJG (eds) (2004) *Plant analysis procedures*. Springer, Netherlands
- Tissue DT, Griffin KL, Turnbull MH, Whitehead D (2001) Canopy position and needle age affect photosynthetic response in field-grown *Pinus radiata* after 5 years of exposure to elevated carbon dioxide partial pressure. *Tree Physiol* 21:915–923. doi:10.1093/treephys/21.12-13.915
- Urban O, Janouš D, Pokorný R et al (2001) Glass domes with adjustable windows: a novel technique for exposing juvenile forest stands to elevated CO<sub>2</sub> concentration. *Photosynthetica* 39:395–401. doi:10.1023/A:1015134427592
- Urban O, Pokorný R, Kalina J, Marek MV (2003) Control mechanisms of photosynthetic capacity under elevated CO<sub>2</sub> concentration: evidence from three experiments with Norway Spruce trees. *Photosynthetica* 41:69–75. doi:10.1023/A:1025808428684
- Valladares F, Niinemets Ü (2008) Shade tolerance, a key plant feature of complex nature and consequences. *Annu Rev Ecol Evol Syst* 39:237–257. doi:10.1146/annurev.ecolsys.39.110707.173506
- Zak DR, Kubiske ME, Pregitzer KS, Burton AJ (2012) Atmospheric CO<sub>2</sub> and O<sub>3</sub> alter competition for soil nitrogen in developing forests. *Glob Change Biol* 18:1480–1488. doi:10.1111/j.1365-2486.2011.02596.x
- Zerbe S (2002) Restoration of natural broad-leaved woodland in Central Europe on sites with coniferous forest plantations. *For Ecol Manag* 167:27–42
- Zuur AF, Ieno EN, Walker NJ et al (2009) *Mixed effects models and extensions in ecology with R*. Springer, New York